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•	INTRODUCTION Eutrophication of surface waters is a major			Spatia es Cre		season	al varia	tions o	of extr	actabl	e P an	d N (m	lg kg⁻	¹) in riv	v <mark>er sedi</mark> n	ments o	of
	environmental issue worldwide (Khan and Ansar,	Location	W-P	STD	MW-P	STD	Olsen P	STD	M3-P	STD	M1-P	STD	TON	STD	NH₄-N	STD	NO ₃ -N
	2005; Smolder et al., 2006). In freshwater	SL1	1.60	0.15	1.47	0.16	8.84	0.64	12.54	1.69	37.02	5.75	222	150	5.55	1.80	1.34
	ecosystems, phosphorus (P) and nitrogen (N) are	SL1 SL2	1.00	0.10		0.10	3.89					4.97	146			1.00	1.43
	commonly considered as main limiting nutrients	SL2 SL3	2.58			0.16		0.57			17.89	3.15	_	60.74		1.72	1.51
	for phytoplankton growth (Howarth et al., 2002;	SL4	1.55	0.18	0.69	0.10	5.96	3.21	10.08	0.38	11.13	0.70	227	149	7.17	5.44	
	Phlips et al., 1997; Piehlera et al., 2004).	SL5	2.09	0.28	2.53	0.20	10.53	2.41	32.48	2.60	37.32	3.35	314	71.58	7.09	4.14	1.92
•	Sugnanded golid norticles (SSD) from surface	SL5B	1.62	0.13	1.46	0.11	11.34	4.04	16.91	0.62	62.33	6.10	440	96.73	6.67	2.39	2.37
	Suspended solid particles (SSP) from surface	SL6	4.19	0.16	3.81	0.48	15.73	3.75	40.36	4.39	180	25.53	1263	482	10.53	5.08	
	runoff waters carrying nutrients can settle down	Spring	2.32	1.55		1.86	7.25				49.33	42.91	284			2.05	
	onto sediment or being transported downstream,	Summer	2.18	1.31	2.10	1.53	7.80					98.87	314			1.24	2.11
	depending on water dynamic conditions. The	Fall Winter	2.16 1.85	0.87 0.88	1.64 1.42	0.81 0.73	8.34 11.81	4.90 5.73	16.88 21.59		53.63 44.37	62.63 35.53	550 426	620 404		5.16 2.52	
	settlement of SSP is commonly observed in																5.10
	upstream. However, under conditions of strong						; MW-P : 1							· · · · · · · · · · · · · · · · · · ·		ich 3	
	water dynamic or change of redox potential, settled SSP can be re-suspended, and result in release of	extra	actable	e P; MI-	P: Mehl	ch I ext	ractable P	? ; STD:	standard	l deviati	on; TOP	N: total o	organic	nıtrogen	•		
	nutrients (N and P) into overlying water body.		6	.0							20.0 —						1
	Therefore, N and P in sediment could serve as an	nt	_		zzz Sprin	a				nent						T	
	internal source . This process is controlled by many	lime	olution 5	.0 -	Sum	-				edin /ater					Ŧ		
	- · · ·) sec	S	.0 - 🖾	💳 Fall 🚾 Winte	÷r				in s ed w	15.0 -						
	factors such as redox potential, pH, salinity, and	AI) ir	ch 1	.0						(IA+¢ nozi							
	physical stability of river water (Perry et al., 2002).	Fe+/	1ehli 3	.0 -						∕/(Fe / dei	10.0 -						
•	In Florida, nutrient enrichment of the Indian River	f P/(by N						I	of F d by							
	Lagoon (IRL) is a major surface water quality	o tio	acted 5	.0 -	Т					ratio ract€							
	issue. Ten Miles Creek (TMC), a major tributary	ur rat		.0	_⊥ ₅∎				3	Molar as exti	5.0 -			Т			
	discharging nutrient-enriched drainage into North	Mola	as extr t				工丙			Me as			ĪΤ	Т		I	
	Fork of Indian River, delivers some 8.6x10 ⁵ kg of	_		.5							7.0						
	N, 9.1x10 ⁵ kg of P, and 3.6x10 ⁸ kg of suspended	ent	⊆ 3	.0 -					т	nt er	6.0 -			-	_		
	solids (SS) to the IRL estuary each year together	dim	solution 5	.0						dime wat	0.0			E C	3		
	with other two artificial canals (Graves and Strom,		0s 2 8	.5 -						n sec cted	5.0 -		Ŧ		_	T	
	1992).	i (IA-	2 lich	.0 -						AI) in nstruc	4.0						
		\sim	Meh							(Fe+Al) reconst			H				
•	The Indian River watershed is a dominant citrus	sf P∕(∫ fq T	.5 -						f P/(l by re	3.0 -			Ţ	Т		
	and vegetable production area in Florida. During	ratio c	extracted by 0 1 1	.0 -							2.0	_ T T	T		ТТ	TTT	
	heavy irrigation and frequent rainfalls, surface	ar ra	extra	F			тΤ	T_		olar ratio o extracted							
	runoff water carries large quantities of soil	Mol	as ex 0	.5 -		_ ∎				/olar Is ext	1.0 -						

particles containing N and P into creeks and canals, and the particles precipitate as part of river sediments or being carried into the Indian **River Lagoon.** To date, few studies have examined the sediments of the TMC as a possible source of N and P and the effects on eutrophication of the estuary.

OBJECTIVES

The objectives of this study were: 1) To determine temporal and spatial variation and availability of N and P in the sediments of TMC, and 2) To examine the possibility of sediment N, P as an internal source of N, P for overlying water body.

MATERIALS AND METHODS

Sediments were seasonally collected from Ten Mile **Creek. Sediment pH and EC were measured using** a pH/ion/conductivity meter. TOC and TON were measured using a Vario MAX C/N analyzer. Available N was extracted by 2M KCl. Available P was determined by the 0.5 M NaHCO₃ (Kuo, **1996).** Labile P, Fe, Ca, and Al were determined by extracting the samples with deionized water, Mehlich 1 (Mehlich, 1953), and Mehlich 3 solution (Mehlich, 1984), respectively. Water extractable NO₃, SO₄, and PO₄ were determined using an IC after the samples were filtered through a 0.45 µm membrane. Mimic river waters were also used to extract sediment samples from Sites SL1-SL4, and mimic sea water for sediment samples from Sites SL5-SL6 where tidal sea water can reach. The concentrations of P, Fe, and Al in the extracts were determined using ICP-AES. Reactive P in water samples was determined by the molybdenum-blue method; total P was determined with unfiltered samples after being digested with acidified ammonium persulfate.

	as exilacted by Ivierificit 1 solution	600.0
lution		500.0
		400.0
v Mehli		300.0
cted p		200.0
as extra		100.0
u	5	500
	as extracted by Mehlich 3 solution	400
(4 (300
		200
		100
-		C

River Sediment—as a Sink/Source of Phosphorus and Nitrogen for Overlying Water Body?

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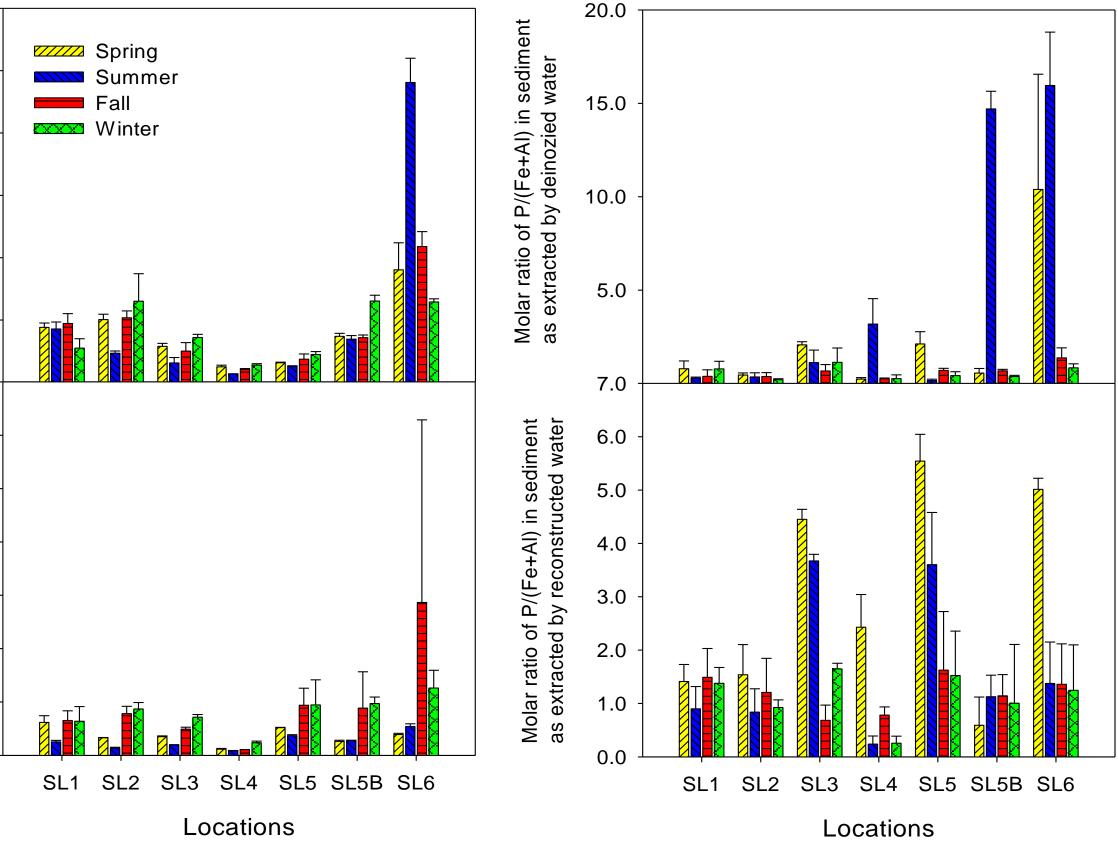


Fig. 1. Ratios of P/(Fe+Al) as measured by different extractants in river sediments of Ten Miles Creek

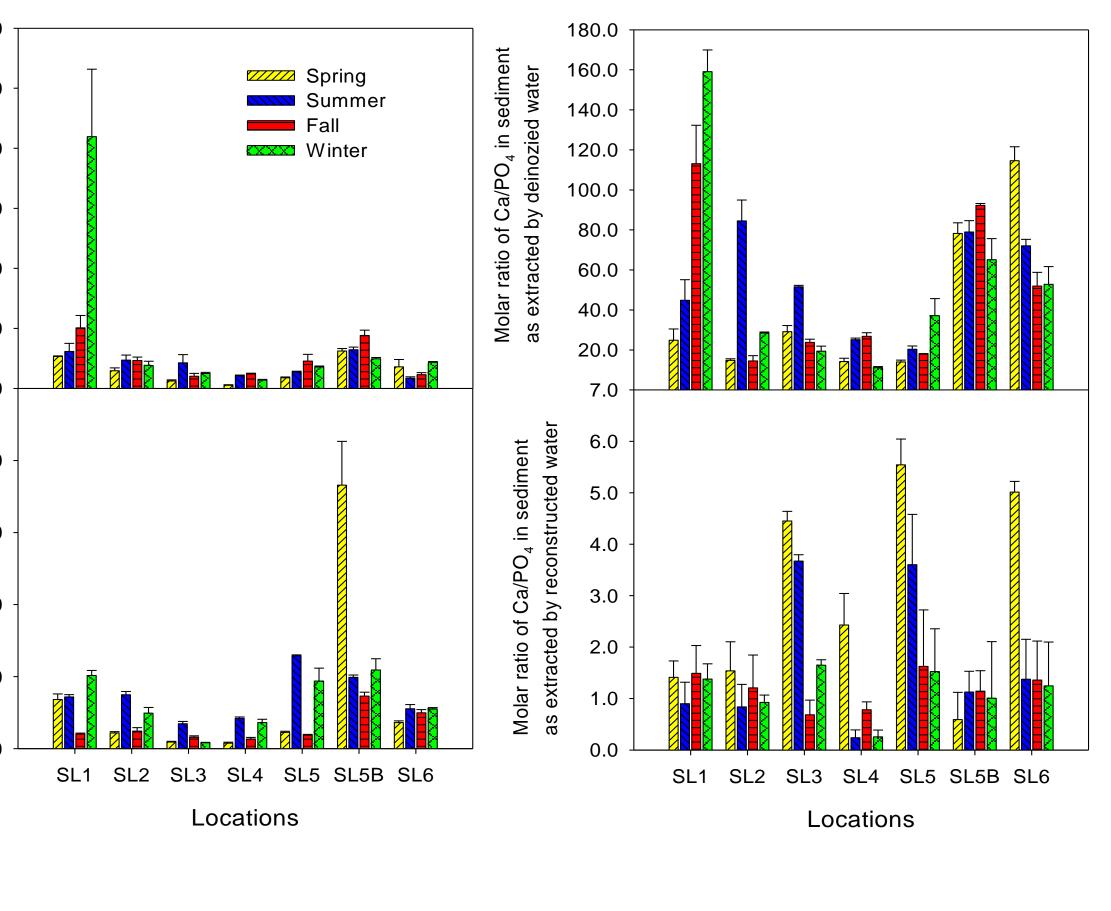


Fig. 2 Ratios of Ca/SO_4 as measured by different extractants in river sediments of Ten Miles Creek

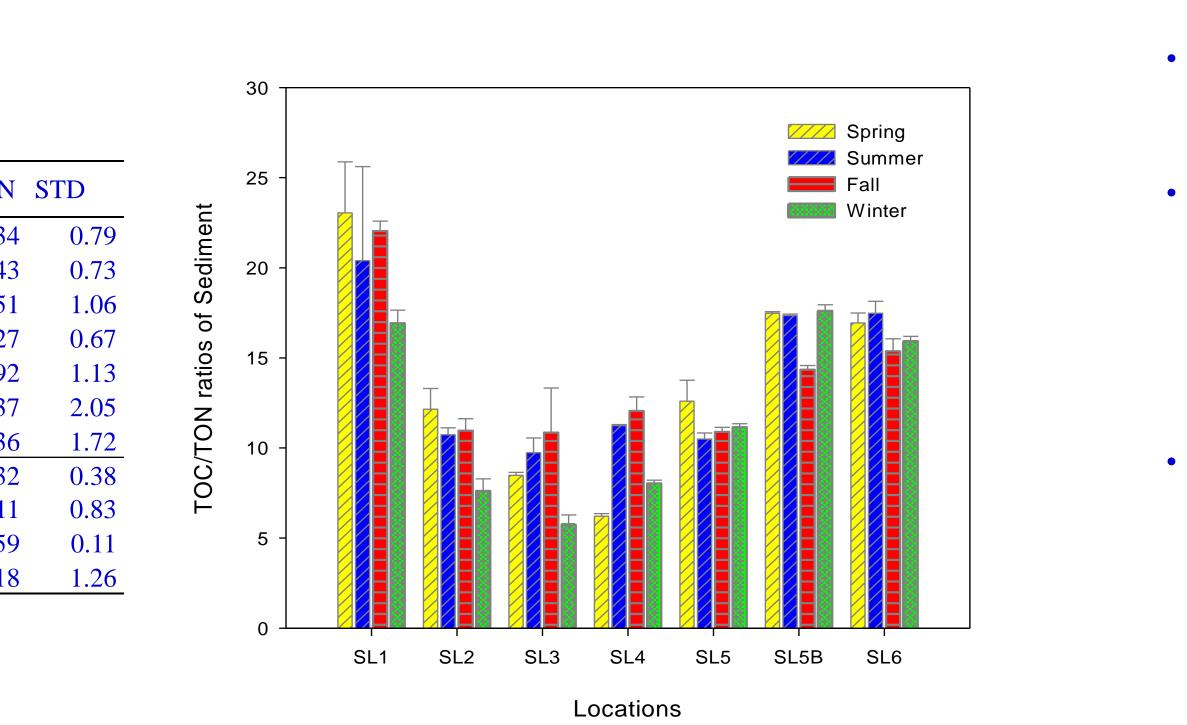


Fig. 3. Ratios of TOC/TON of different seasons in river sediments of Ten Miles Creek.

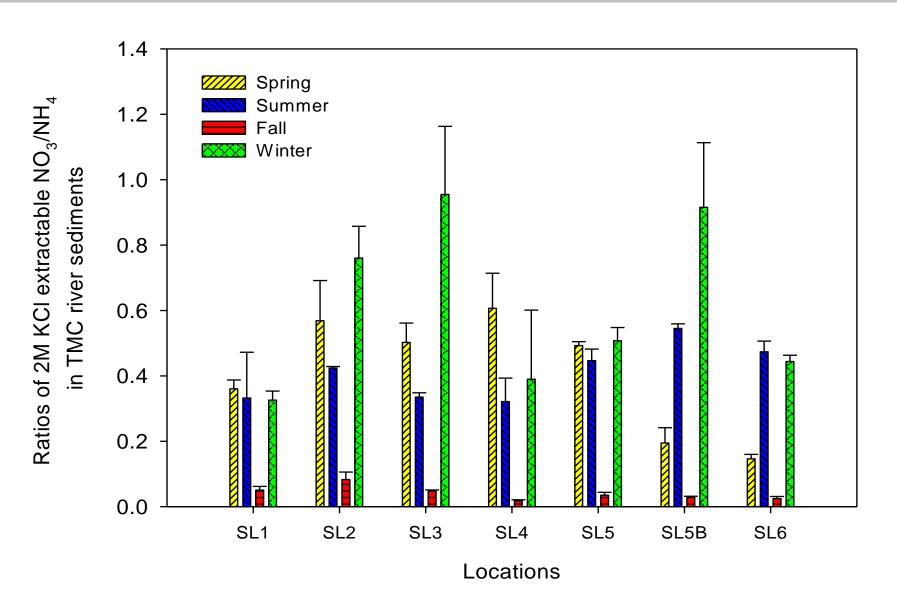


Fig. 4. Ratios of 2M KCl extractable NO₃/NH₄ of different seasons in river sediments of Ten Miles Creek .

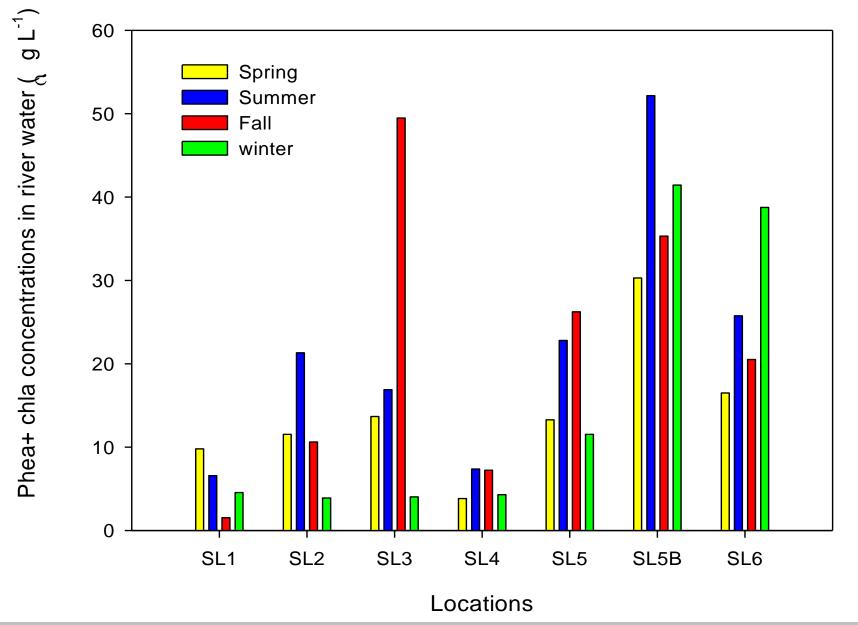


Fig. 5. Concentrations of phea (pheophytin)+chla (chlorophyll a) in river waters of overlying water bodies in the TMC.

Table 2. Correlations between ratios of M1 and M 3 extractable
 P/(Fe+Al) and water extractable PO_4 -P, dissolved total P (DTP), the ratios of NO₃/SO₄, 2M KCl extractable ratio of NO₃/NH₄ in TMC sediments

Ratios	DI water extractable DTP	Mimic water extractable DTP	DI water extractable PO ₄ -P	DI water extractable NO ₃ /SO ₄	KCl extractable NO ₃ /NH ₄
P/(Fe+Al) _{M1}	0.686**	0.614**	0.232*	-0.261	-0.105
P/(Fe+Al) _{M3}	0.519**	0.547**	0.448**	-0.174	-0.020
(Ca/PO ₄) _{M1}	-0.080	-0.085	-0.164	-0.153	-0.108
$(Ca/PO_4)_{M3}$	0.267*	0.236 *	-0.109	-0.334 *	-0.197

Statistic sample number=84; ratios refer to molar ratios.

• Total dissolved P was determined using ICP-AES after filtered through a 0.45 µm membrane. The concentrations of NO₃-N, PO₄-P, and SO₄-S in water were determined using IC.

Chlorophyll in water collected by filtering was extracted using 95% ethanol in a 78°C water bath for 5 min. **Chlorophyll a and pheophytin concentrations in supernatant** solution were colorimetrically determined using a dual beam spectrophotometer

RESULTS AND DISCUSSION

Extractable P in river sediment varied greatly in the range of 0.69-180 mg kg⁻¹, and changed among the different extractions in the order of W-P) ≈ W-P < Olsen P< M3-P< M1-P (Table 1). The most downstream location (SL6) had the highest extractable P. Winter had the lowest W-P, RW-P, and M1-P, but the highest Olsen-P and M3-P, while spring had the highest W-P, MW-P, but lowest Olsen-P (Table 1). Total organic N (TON) varied from 143- 1263 mg kg⁻¹ with peak occurred at SL6 (Table 1). KCl extractable NH₄-N and NO₃-N were in the range of 3.51-10.5 and 1.27-2.37 mg kg⁻¹ with highest values occurred at the most downstream location **(Table 1).**

M1, M3 and DI water extractable P/(Fe+Al) ratios were high at Sites SL5, SL5B, and SL6. Summer had high ratios in downstream locations, suggesting high P availability in downstream sediments (Fig. 2). The ratios of Ca/PO4 were less than 100, but larger than 5/3 [atomic ratios of Ca₅(PO₄)₃(OH, F, Cl)], suggesting an over-saturation of Ca to **PO₄** in the river sediment (Fig. 3). TOC/TON ratios were in the range of 5.74-23 varying with location and seasons (Fig.

Ratios of KCl extractable NO_3/NH_4 were less than 1, and the lowest ratios were found in fall, highest in winter, indicating a relatively reducing environment in the sediments (Fig. 3).

• Molar ratios of P/(Fe+Al) in M1 and M3 extraction were positively correlated with DI water extractable PO₄-P, DTP, and mimic water extractable DTP at significant levels of P<0.01(Table 2). However, molar ratios of Ca/PO₄ in M3 extraction showed positive correlations (P<0.05) with DI water and mimic water extractable DTP, but had a negative correlation with DI water extractable NO₃/SO₄ (Table 2). These correlations showed that the availability of P and N in TMC sediment was more influenced by chemical adsorptiondesorption process than redox condition .Water hydrological background might have some constraints on P and N availability.

Phytoplankton was found blooming at downstream locations, which is related to high concentrations of N and P in underlying river sediments (Fig. 5).

CONCLUSIONS

. The concentrations of extractable N and P in TMC river sediment are high, especially at downstream location, and mainly controlled by chemical adsorption-desorption process.

2. River sediments in the TMC, especially at its downstream could serve as a potential secondary P or N source for overlying river water, and probably be a driver to the blooming of phytoplankton.

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