

Introduction

- In the U.S. Midwest, many states, including Iowa, Ohio, Illinois and Minnesota, are developing strategies to reduce export of nitrate-N (NO₃-N) from agricultural nonpoint sources to the Mississippi River.
- Management of soil organic carbon (SOC) and nitrogen (N) is considered a key component to prevent N loss to groundwater and artificial drainage.

Results





• NO₃-N is being lost from the Rapid Creek agricultural row crop system. The goal of this study was to identify spatial patterns and evaluate the water quality risks associated with the surficial soils in the subbasin.

SOC concentrations exhibited no overt spatial patterns in the watershed.

- It has been estimated that 24 to 89% of SOC has been lost from the North American prairie since settlement. Improving SOC stocks can limit nutrient loss and improve water quality, improve the economic potential of cropland, and provide resilience needed for global climate change and food security.
- Despite inherent challenges in quantifying SOC and N in agricultural landscapes, enhancing and manipulating SOC and N pools offers potential to increase crop production without sacrificing water quality.
- Our study objectives were to 1) evaluate SOC, N and C/N ratios in the top 20 cm (plow layer) of the Rapid Creek subbasin; 2) quantify watershed-scale storage of SOC and N based on landscape positions and slope class criteria; and 3) assess the groundwater quality risks associated with the shallow soils in the subbasin.

Methods

• The 780 ha subbasin is located in the northern portion of the 88 km² HUC 12 Rapid Creek watershed in Johnson County, Iowa. The subbasin is situated on the southern edge of the Iowan Surface landform region in eastern Iowa (Figure 1).

- Table 1. Summary of soil texture and nutrient SOC and N concentrations varied across the concentrations for 0-20 cm deep soils in the Rapid Creek subbasin study area. subbasin (Figure 2).
- N concentrations ranged from 132 to 1031 kg ha⁻¹ and averaged 493 \pm 195 kg ha⁻¹ in the 51 samples (Table 1). However, N concentrations were biased by a single sample with a concentration approximately an order of magnitude higher than the other samples (1031 kg ha⁻¹). Without this sample, N concentrations averaged 482 kg ha⁻¹ and the standard deviation decreased
- SOC concentrations ranged from 3 to 12 Mg ha⁻¹ and averaged 8 \pm 2 Mg ha⁻¹ (Table 1).
- C/N ratios averaged 18 ± 6 for the individual samples (Table 1). The C/N ratio based on the subbasin SOC and was 16.

to 182.

• SOC and N concentrations were distributed (p<0.05) whereas C/ were not (p>0.05). Minor variation

• 3-6

| | | SOC | | C/N ratio | | Ν | |
|--------------------|-----------|-----------|------|-----------|------|-----------|--------|
| | | (Mg ha⁻¹) | | | | (Kg ha⁻¹) | |
| Landscape position | | | | | | | |
| n=3 | Headland | 9a | +/-2 | 15 | +/-4 | 626 | +/-817 |
| n=6 | Shoulder | 6b | +/-3 | 22 | +/-3 | 296 | +/-578 |
| n=34 | Backslope | 8a | +/-2 | 17 | +/-1 | 805 | +/-243 |
| n=8 | Toeslope | 7ab | +/-3 | 19 | +/-2 | 426 | +/-501 |
| Slope | class | | | | | | |
| n=4 | А | 8 | +/-1 | 23 | +/-3 | 371 | +/-99 |
| n=27 | В | 8 | +/-1 | 18 | +/-1 | 516 | +/-38 |
| n=17 | С | 8 | +/-1 | 16 | +/-1 | 506 | +/-48 |
| n=3 | D | 7 | +/-1 | 22 | +/-4 | 377 | +/-114 |

Table 2. Comparison of mean soil organic carbon (SOC) and nitrogen (N) stocks, and C/N ratios organized by landscape position and slope class. Letters report significance (p<0.05).

| • | | | | | | | |
|----------------------|-------|----------|----------|-------------|----------|----------|-----------------|
| N averages | Slope | SOC | N | Soil series | SOC | Ν | series classifi |
| | Α | 1.49E+03 | 6.91E+01 | Ackmore | 3.28E+02 | 2.23E+01 | schemes prov |
| | В | 3.01E+03 | 1.94E+02 | Colo-Ely | 8.85E+02 | 8.10E+01 | (Table 3). N st |
| normally N ratios | С | 1.58E+03 | 1.00E+02 | Downs | 6.42E+02 | 3.53E+01 | compared to |
| ons of SOC | D | 1.30E+02 | 7.02E+00 | Fayette | 3.26E+02 | 1.85E+01 | between qua |

Spatially, nutrient concentrations varied across transects and within individual fields (Figure 2).

We found higher concentrations of SOC and N on the backslope landscape positions while the lowest concentrations were found on shoulders (Table 2).

No significant differences were identified when SOC comparisons were made by soil slope class. No significant differences were identified when comparing C/N ratios or N among landscape positions or slope class (Table2).

 We estimated the total mass of SOC and N present in the subbasin using bulk density measurements and both slope class and soil ications. Both classification vided similar estimates of SOC stocks were higher when ing soil series classifications

slope class. Overall, the RPD antification methods was

Some fields sampled for our study have been previously enrolled in CRP. However, we did not observe any association of higher SOC in soils with CRP history.

• Sampling data indicated that only unstable shoulder positions showed some evidence for lower SOC levels. The lack of significant spatial variability is likely due to the effects of long- and short-term soil erosion processes.



N concentrations did not show any systematic variations that would suggest potential source areas for NO₃-N loss which suggests that management history and landscape position are not major controls on N within the subbasin.



- Figure 1. Location map of the Iowa Surface landform region (A) and the Rapid Creek subbasin (B).
- Soils are primarily formed in loess and alluvium.
- A 500 x 200 m sampling grid was created for the subbasin. In all, 51 sites were sampled along 8 transects. All of the sample sites were located in agricultural row crop fields where tillage consisted of either mulch or no-till (Figure 1).
- Soil samples, 20 cm deep, were collected at each location approximately 30 days after planting in 2015.
- Landscape positions were determined at the time of sampling and





Figure 2. Spatial nutrient graphs showing distribution of (A) soil organic carbon (SOC) (B) total nitrogen (N) and (C) C/N ratio.



- In our study, C/N ratio, like SOC and N, was not consistently variable and could not be used to identify areas of greater N leaching threat.
- Identifying stocks of SOC and N is required to better understand the soils ability to buffer the effects of climate change and improve water quality. The similarities of SOC stocks between slope class and soil series provide evidence that both are appropriate for estimating total SOC stocks at a watershed scale.

Implications

- Study results indicate that 1) there is no "hot spot" identifying a row crop field or management practice contributing to greater NO₃-N leaching risk, and 2) a basin-wide solution is needed to reduce NO3-N loss.
- Results from our study emphasize the potentially beneficial use of cover crops or living mulch in the Rapid Creek subbasin to reduce NO₃-N export. Adding C sources would increase the basin-wide soil C/N ratios and reduce the potential for spring nitrification when NO₃-N concentrations in the stream exceed 10 mg l⁻¹. Moreover, the cover crops provide late season N to the soil microbial

included headland, shoulder, backslope, and toeslope.

- Bulk density was determined at 2 depths (5-10 cm and 15-20 cm) and averaged.
- SOC and total N were determined by elemental analysis via dry combustion.
- Total mass of SOC and N was estimated based on percent by mass and mean bulk density of the plow layer for each site. These data were used to estimate total mass of each soil property for the entire subbasin assuming equal mass of each soil property for like landscape positions and slope class.

Figure 3. Probability plots for soil organic carbon (SOC), total nitrogen (N), and C/N ratio.

community and a slow feed of N to the crop. Over time, adding cover crops or other carbon sources serves to build up the SOC pool and increase C/N ratios, thereby lessening potential NO₃-N loss from the row crop agroecosystem.

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